Author's Accepted Manuscript

Development of analytical methodologies to assess recalcitrant pesticide bioremediation in biobeds at laboratory scale

Anisleidy Rivero, Silvina Niell, M. Pía Cerdeiras, Horacio Heinzen, María Verónica Cesio



 PII:
 S0039-9140(16)30094-7

 DOI:
 http://dx.doi.org/10.1016/j.talanta.2016.02.025

 Reference:
 TAL16343

To appear in: Talanta

Received date:17 November 2015Revised date:10 February 2016Accepted date:10 February 2016

Cite this article as: Anisleidy Rivero, Silvina Niell, M. Pía Cerdeiras, Horacia Heinzen and María Verónica Cesio, Development of analytical methodologies ta assess recalcitrant pesticide bioremediation in biobeds at laboratory scale *Talanta*, http://dx.doi.org/10.1016/j.talanta.2016.02.025

This is a PDF file of an unedited manuscript that has been accepted fo publication. As a service to our customers we are providing this early version o the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting galley proof before it is published in its final citable form Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain

1	Development of analytical methodologies to assess recalcitrant pesticide
2	bioremediation in biobeds at laboratory scale
3	
4	Anisleidy Rivero ^{a,c} , Silvina Niell ^d , M. Pía Cerdeiras ^b , Horacio Heinzen ^{a,d} , María
5	Verónica Cesio ^{a*}
6	a. Cátedra de Farmacognosia, Facultad de Química, Universidad de la República, Gral.
7	Flores 2124, Montevideo, CP 11800, Uruguay. heinzen@fq.edu.uy, cs@fq.edu.uy.
8	b. Cátedra de Microbiología, Facultad de Química, Universidad de la República, Gral.
9	Flores 2124, Montevideo, CP 11800, Uruguay. mcerdeir@fq.edu.uy
10	c. Laboratorio de Microbiología, Laboratorio Tecnológico del Uruguay (LATU), Parque
11	Industrial Municipal Barrio Anglo, Fray Bentos, CP 65 000, Uruguay.
12	arivero@latu.org.uy
13	d. Polo Agroalimentario y Agroindustrial, Departamento de Química del Litoral,
14	CENUR Litoral Norte Sede Paysandú, Universidad de la República, Ruta 3 km 363,
15	Paysandú, CP 60 000, Uruguay. sniell@cup.edu.uy
16	*corresponding author: María Verónica Cesio ^a ; cs@fq.edu.uy.
17	

19 ABSTRACT:

To assess recalcitrant pesticide bioremediation it is necessary to gradually increase the 20 complexity of the biological system used in order to design an effective biobed 21 assembly. Each step towards this effective biobed design needs a suitable, validated 22 analytical methodology that allows a correct evaluation of the dissipation and 23 24 bioconvertion. Low recovery yielding methods could give a false idea of a successful biodegradation process. To address this situation, different methods were developed and 25 validated for the simultaneous determination of endosulfan, its main three metabolites, 26 and chlorpyrifos in increasingly complex matrices where the bioconvertor 27 basidiomycete Abortiporus biennis could grow. The matrices were culture media, bran, 28 29 and finally a laboratory biomix composed of bran, peat and soil. The methodology for the analysis of the first evaluated matrix has already been reported. The methodologies 30 developed for the other two systems are presented in this work. The targeted analytes 31 were extracted from fungi growing over bran in semisolid media YNB (Yeast Nitrogen 32 Based) with acetonitrile using shaker assisted extraction. The salting-out step was 33 performed with MgSO₄ and NaCl, and the extracts analyzed by GC-ECD. The best 34 methodology was fully validated for all the evaluated analytes at 1 and 25 mg kg⁻¹ 35 yielding recoveries between 72 and 109% and RSDs <11% in all cases. The application 36 of this methodology proved that A. biennis is able to dissipate 94 % of endosulfan and 37 87 % of chlorpyrifos after 90 days. Having assessed that A. biennis growing over bran 38 can metabolize the studied pesticides, the next step faced was the development and 39 40 validation of an analytical procedure to evaluate the analytes in a laboratory scale biobed composed of 50 % of bran, 25 % of peat and 25 % of soil together with fungal 41 micelium. From the different procedures assayed, only ultrasound assisted extraction 42 43 with ethyl acetate allowed recoveries between 80 %-110 % with RSDs <18 %.

Linearity, recovery, precision, matrix effect and LODs/LOQs of each method were studied for all the analytes: endosulfan isomers ($\alpha \& \beta$) and its metabolites (endosulfan sulfate, ether and diol) as well as for chlorpyrifos. In the first laboratory evaluation of these biobeds endosulfan was bioconverted up to 87 % and chlorpyrifos more than 79 % after 27 days.

- Keywords: Basidiomycete; Recalcitrant pesticides; Bioremediation; Method
 validation

54 1. Introduction

Environmental pollution caused by hazardous wastes containing recalcitrant 55 xenobiotic chemicals has become a major problem that threatens the sustainability of 56 the ecosystems as well as human health. Unlike the naturally occurring organic 57 compounds that are readily degraded upon introduction into the environment, some of 58 these synthetic chemicals are extremely resistant to biodegradation by native 59 microorganisms. Particularly, surface and ground waters are exposed to diffuse 60 pollution due to pesticides via percolation, runoff, drainage and drift (i.e. movement of 61 62 airborne spray droplets) as well as spills during equipment washing. These critical points are the main sources of soil and water contamination, contributing significantly 63 to the deterioration of natural water sources [1, 2-5]. This is an issue of global strategic 64 importance as groundwater represents about 98 % of the available fresh water of our 65 planet [6]. Several field surveys and measurement campaigns on a catchment scale have 66 demonstrated that 40–90 % of surface water contamination by pesticides is attributable 67 to direct losses of the active ingredients [1, 3-5]. To protect ground and surface water 68 quality, practical solutions to minimize the pesticide entrance into hydrographic 69 networks are continuously investigated. One of the possible strategies to minimize it is 70 to degrade the pesticides before releasing the machinery washings and container triple 71 rinses to the environment. 72

As the degradation of recalcitrant pesticides occurs slowly under natural conditions,the process has to be enhanced.

A newly explored route to diminish pesticide spillage into waters or in places where it can lixiviate to ground water is the biobeds technology. Biobeds are environmentally friendly solutions for the remediation of impacted zones as the pesticides convertor and degrading agents are saprophytes microorganisms. This methodology has been used in

Europe for several years and it is now being evaluated in South America as a friendly 79 alternative for remediation of recalcitrant compounds [7, 8]. Ligninolitic fungi are 80 biodegradetors of particular interest because they have shown to degrade and mineralize 81 a large variety of recalcitrant compounds due to the nonspecificity of their enzyme 82 machinery [9, 10]. White rot fungi produce a number of extracellular oxidative enzymes 83 including laccases, lignin peroxidases and manganese peroxidases, which are normally 84 involved in the breakdown of the plant structural material lignin [11] and are also 85 responsible for the degradation of xenobiotics. 86

Challenging targets to assess biobeds suitability are organochlorine (OC) 87 agrochemicals. They are persistent compounds which have been accumulating in the 88 biosphere after decades of massive application in agriculture. Among them, Endosulfan, 89 is an actually banned OC which had been used in many countries until recent years, 90 91 leaving highly contaminated zones as an unsolved problem. In addition, there are also huge amounts of already synthesized endosulfan that have to be stored, waiting for a 92 93 final destination. At the same time, massive use of highly toxic endosulfan, has been substituted by chlorpyrifos situation that renders an even more complicated scenario. 94

The proof of the efficacy of a bioremediation process is sustained on validated, 95 highly sensitive analytical methods that assess the efficiency of the whole process. 96 Agrochemical biodegradation has to be assayed in the laboratory, with microorganisms 97 growing in conventional culture media, in order to select the microorganisms capable of 98 dissipating them. Once the microorganisms have been selected and the efficiency of the 99 transformation process evaluated, trial biobeds made of soil, peat and bran are assayed 100 at a lab scale. Dissipation kinetics within the bioreactor is established, and the residual 101 102 water that lixiviates through it is evaluated for pesticide non detection [12].

In a previous communication our group described ligninolitic fungi capable of degrading recalcitrant pesticides using endosulfan as a model compound in culture media, highlighting the importance of having analytical methods that can assure the effective dissipation of the pesticides into harmless products [13]. This work presents the advances in the development of a bioreactor using endosulfan and chlorpyrifos as model compounds, based in the Swedish biobed design [12] and native Basidiomycetes fungi as bioconvertors.

For each of the three different complex matrices employed: culture media, bran and 110 the biobmix where the fungal mycelium grows and the biotransformation is performed, 111 different methodologies were developed. Linearity, recovery, precision, matrix effect 112 and LODs/LOQs, were determined for each of the endosulfan isomers ($\alpha \& \beta$) and some 113 of its metabolites (endosulfan sulfate, endosulfan ether and diol) as well as for 114 115 chlorpyrifos. Although chloryrifos metabolites, particularly TCP (3,5,6trichloropyridinol) [14], is toxicologically important from an environmental point of 116 117 view [15], only the dissipation of the parent compound as a model has been followed in 118 this study. At this point, our interest was focused in the improvement of the biobed design and future work will need to study TCP degradation in the final biobed. 119

120

- 121 2. Materials and methods
- 122
- 123 2.1. Standards and reagents

124

Analytical grade organic solvents, pesticide residues free were purchased from Merck
(Darmstandt, Germany). Pesticide standards and the internal standard were from Dr.
Ehrenstorfer (Augsburg, Germany, 99 %). The culture media were provided by Difco.

128	Bran and land peat were commercially available. Magnesium Sulfate and Sodium
129	Sulfate p.p.a from Sigma-Aldrich (St.Louis, MO, USA); Sodium Chloride from Merck
130	(Darmstandt, Germany), Silica: MN Kiesel 60 from J.T. Baker (State of Mexico).
131	Stock solutions were prepared from the standard substances at 1000 and 2000 mg L^{-1} in
132	ethyl acetate. Working standard mixtures were prepared by appropriately diluting the
133	stock solutions with ethyl acetate. All solutions were stored at 4°C.
134	
135	2.2 Apparatus and experimental conditions.
136	
137	Gas chromatographic (GC) analyses were performed using a Shimadzu GC 17A
138	equipped with an ECD detector and a PTV injector using internal standard method. All
139	compounds were resolved on a capillary column Mega 68 (30m, 0.32 mm ID, 0.25 μm
140	film thickness) Mega Legnano (Italy). The experimental conditions were as follow:
141	PTV, 60°C (0.3min), then 5°C min ⁻¹ to 280°C (40min). Oven temperature, 100°C
142	(3min), 100-180°C at 10°C min ⁻¹ , 180°C (15min), then 180-270°C at 5°C min ⁻¹ , 270°C
143	(10 min). Detector temperature, 280°C.
144	Orbital shaker: SL1 Stuart (Staffordshire,UK). Ultrasonic bath:
145	Elma®Transsonic T460/H. Centrifuge IEC: HNS-II (U.S.A.).
146	
147	2.3 Microbiological matrix preparation for validation study
148	Y .
149	2.3.1 Matrix A (Bran)
150	The prepared inoculum was added to a mixture of 5 mL of semisolid medium
151	YNB (Yeast Nitrogen Base) and 2.2 g of bran which was previously homogenized in a
152	water bath for 90 min at 45°C. Then, it was incubated at 28±2°C for 20 days.

153	Inoculum: A portion of 1 cm diameter of the external growth of preinoculum was added
154	to 10 mL of Malt Extract medium and incubated at 28±2 °C for 10 days.
155	Preinoculum: Abortiporus biennis was cultured in solid media Potato Dextrose Agar
156	(PDA) at 28±2 °C for 5 days.
157	
158	2.3.2 Matrix B (Biobed)
159	The laboratory scale biobed was prepared by adding 20 times matrix A to 1 kg
160	of a mixture containing 50 % bran, 25 % soil and 25 % peat and incubated for 30 days
161	at room temperature.
162	
163	2.4 Recovery tests
164	
165	2.4.1 Recovery tests for matrix A
166	For recovery studies the matrix was prepared spiking at two levels: 1 and 25 mg
167	kg ⁻¹ with chlorpyrifos, endosulfan and its metabolites. Levels 1 and 25 mg kg ⁻¹ were
168	prepared in quintuplicate by adding 0.3 and 3.5 mL respectively of a solution containing
169	100 mg L ⁻¹ of endosulfan α and β , endosulfan ether, endosulfan diol, endosulfan sulfate
170	and chlorpyrifos to matrix A.
171	
172	2.4.2 Recovery tests for matrix B
173	For recovery studies the matrix was prepared spiking at three levels: 1, 25 and
174	50 mg kg ⁻¹ with endosulfan and its metabolites and 1, 30 and 60 mg kg ⁻¹ with
175	chlorpyrifos. Level 1 mg kg ⁻¹ was prepared in quintuplicate by adding 0.05 mL of a
176	solution containing 200 mg L^{-1} of chlorpyrifos, endosulfan α and β , endosulfan ether,
177	endosulfan diol and endosulfan sulfate to grown fungi in 10 g of matrix B (laboratory

178	biobed); level 25 and 50 mg L^{-1} for endosulfan and its metabolites were prepared by
179	adding 0.12 mL and 0.25 mL respectively of a 2025 mg L^{-1} solution and level 25 and 60
180	mg L ⁻¹ for chlorpryrifos were prepared by adding 0.12 mL and 0.30 mL respectively of
181	a 2020 mg L^{-1} solution.
182	
183	2.5 Extraction and clean up methods comparison
184	
185	2.5.1 Bran (Matrix A)
186	The evaluated methods were:
187	1. Extraction with ethyl acetate and homogenization in Stomacher® and clean up using
188	open column packed with silica and sodium sulfate.
189	2. Extraction with ethyl acetate and homogenization in orbital shaker for 24 h and clean
190	up using open column packed with silica and sodium sulfate.
191	3. Extraction with acetonitrile and water, homogenization in orbital shaker for 2 h and
192	salting out with NaCl and MgSO ₄ .
193	The comparison of the recovery results obtained with the three extraction
194	methods tested was performed spiking at level 50 mg L^{-1} of endosulfan and 60 mg L^{-1}
195	of chlorpyrifos.
196	2.5.1.1. General aspects for all the extraction procedures for Matrix A:
197	The internal standard (IS) used was methyl bromophos, the final aliquot was
198	evaporated with a gentle stream of nitrogen, redissolved with IS and diluted to volume
199	with ethyl acetate to obtain a final vial concentration of 1 mg L^{-1} of the IS and the
200	evaluated. The chromatographic analysis was performed in a GC-ECD under the
201	conditions explained in "Apparatus and experimental conditions".
202	

	ACCEPTED MANUSCRIPT				
203	Extraction and clean up methods for matrix A are summarized in table 1 so that they can				
204	be easily compared.				
205					
206	Table 1.				
207					
208					
209	2.5.2 Laboratory scale Biomix (Matrix B)				
210	The compared methods were:				
211	1. Extraction with acetonitrile and water, homogenization in orbital shaker and salting-				
212	out with magnesium sulfate and sodium chloride.				
213	2. Extraction with ethyl acetate and homogenization in ultrasonic bath (ultras) for 15				
214	min.				
215	3. Extraction with acetonitrile and homogenization in ultrasonic bath for 15 min.				
216	4. Extraction with ethyl acetate and homogenization in shaker with iron pellets				
217	assistance.				
218	5. Extraction with ethyl acetate and homogenization in ultrasonic bath (3 cycles of 15				
219	min. each one).				
220	6. Extraction with ethyl acetate in soxhlet apparatus.				
221					
222	2.5.2.1 General aspects for all the extraction procedures for Matrix B:				
223	Sample humidity was adjusted with a solution of NaCl (10 %). The internal				
224	standard (IS) used was methyl bromophos. The final aliquot was evaporated with a				
225	gentle stream of nitrogen and redissolved with a solution of the IS in ethyl acetate and				
226	diluted to volume with ethyl acetate in a 25 mL volumetric flask obtaining a final vial				

227	concentration of 1 mg L ⁻¹ of the IS. The chromatographic analysis was performed in a
228	GC-ECD under the conditions explained in "Apparatus and experimental conditions".
229	Extraction methods for matrix B are summarized in table 2 so that they can be easily
230	compared.
231	Table 2.
232	(*) Addition of 1 iron pellet per g of sample.
233	(**) 3 cycles of 15 min each one and manual agitation between cycles.
234	
235	2.6 Sample preparation for endosulfan/its metabolites and chlorpyrifos analysis.
236	
237	2.6.1 Final method for Bran (Matrix A)
238	
239	Matrix A was finally analyzed using method 3 summarized under "Extraction
240	and clean up methods comparison, Bran (Matrix A)":
241	The whole flask of rice straw with the fungi grown in semisolid media was extracted
242	with 15 mL acetonitrile and 15 mL water in an orbital shaker apparatus for two hours.
243	Then, 20g of $MgSO_4$ and 2g NaCl were added and placed in an orbital shaker for 5
244	hours. Finally it was centrifuged 10 min at 3000 rpm and an aliquot of 0.5 mL of the
245	extract was evaporated with a gentle stream of nitrogen, redissolved with IS and diluted
246	to volume with ethyl acetate to obtain a final vial concentration of 1 mg L^{-1} of the IS
247	and analytes evaluated. The chromatographic analysis was performed in a GC-ECD
248	under the conditions described in "Apparatus and experimental conditions".
249	
250	2.6.2 Final Method for the laboratory scale Biomix (Matrix B)

Matrix B was finally analyzed using method 5 summarized in "Laboratory scale
Biomix analysis (Matrix B)":

254

255	10 g of biobed mix and inoculum (Microbiological matrix preparation for
256	validation study: Matrix B (Biobed)) were placed in a 250 mL Erlenmeyer flask and 10
257	mL solution of 10 % NaCl and 50 mL of ethyl acetate were added and extracted in an
258	ultrasonic bath (3 cycles of 15 min each one). The final aliquot (0.5 mL) was evaporated
259	with a gentle stream of nitrogen and redissolved with a solution of the IS in ethyl
260	acetate and diluted to volume with ethyl acetate in a 25 mL volumetric flask obtaining a
261	final vial concentration of 1 mg L^{-1} of the IS and analytes evaluated. The
262	chromatographic analysis was performed in a GC-ECD under the conditions explained
263	in "Apparatus and experimental conditions".
264	
265	2.7. Methods validation
266	2.7.1 Matrix A
267	The following parameters were evaluated for the analytical method validation:
268	linearity, recoveries, repeatability (RSDr), within-laboratory reproducibility (RSDwR),
269	LOQs and LODs. All the tests were performed at two levels, five replicates in three
269 270	LOQs and LODs. All the tests were performed at two levels, five replicates in three different days. Solvent and matrix matched calibration curves were compared and
269 270 271	LOQs and LODs. All the tests were performed at two levels, five replicates in three different days. Solvent and matrix matched calibration curves were compared and matrix effects were quantified. Matrix effects were evaluated at two different amounts
269 270 271 272	LOQs and LODs. All the tests were performed at two levels, five replicates in three different days. Solvent and matrix matched calibration curves were compared and matrix effects were quantified. Matrix effects were evaluated at two different amounts of matrix 0.5 and 0.02 g mL ⁻¹ of extract which correspond to the dilutions needed to
269 270 271 272 273	LOQs and LODs. All the tests were performed at two levels, five replicates in three different days. Solvent and matrix matched calibration curves were compared and matrix effects were quantified. Matrix effects were evaluated at two different amounts of matrix 0.5 and 0.02 g mL ⁻¹ of extract which correspond to the dilutions needed to cover the wide range of concentrations 1 to 25 mg L ⁻¹ .

274 2.7.2 Matrix B

The parameters evaluated for the analytical method validation exposed in the item "Methods validation, Matrix A" were evaluated in the same conditions. Matrix

effects were evaluated at three different amount of matrix 0.2, 0.07 and 0.03 g mL⁻¹ of extract which corresponds to the dilutions needed to cover the wide range of concentrations 1 to 50 and 60 mg L⁻¹ for endosulfan and chlorpyrifos respectively.

280 The different parameters evaluated were calculated as explained below:

RSDr presented are an estimation of the precision of the method. They were calculated as the relative standard deviation (coefficient of variation) of the five replicates from the recovery test measurement of each analyte, obtained using the same method on the same sample in a single laboratory over a short period of time, during which differences in the materials and equipment used and the analysts involved did not occur.

RSDwR presented are within laboratory reproducibility. They were calculated as the relative standard deviation (coefficient of variation) of all the replicates from the recovery test measurement of each analyte obtained using the same method in a single laboratory, by different analysts, and over a period in which differences in the materials and equipment occurred.

LOQs limits of quantitation (quantification) presented are the lowest concentration of the analyte that has been validated with acceptable accuracy (recoveries between 70-120% and RSDs < 20%) by applying the complete analytical method as recommended by DG-SANCO [16].

LODs limits of detection presented are the lowest concentration of the analyte that presents a response with a signal to noise ratio equal or above 3 by applying the complete analytical method.

298

299

300

301 3. Results and Discussion

302

The global strategy followed to accomplish the biobed design at laboratory scale is based in the developing of fit for purpose analytical methodologies for very complex and heterogeneous matrices, in which the pesticides and their metabolites have to be determined.

The Swedish biobed is composed of a mixture of peat, straw and soil (1:2:1). As the selected fungi did not grow properly over straw to yield enough biomass, we looked for other fungi grow promoting substrates. It has been reported that straw can be replaced by other agricultural byproducts with equal or superior biotransformation capacities [17].

Looking for straw alternative substrates, good biomass amounts were obtained with the fungi growing over cereal bran, which is a more nutritive substratum. Therefore, the bioconversion of the targeted pesticides by fungi growing over bran was assayed and the best analytical methodology for this matrix was selected and validated. The pesticides to evaluate the performance of the bioreactor were chosen based on their use in agriculture their persistence and their toxicity [18-20].

318

319 3.1. Methods comparison

320 3.1.1. Bran (Matrix A)

321

Three methods were evaluated aiming to find a good methodology to analyze chlorpyrifos, endosulfan and its metabolites in the microbiological media prepared in the laboratory with bran as the major component of the system. The combination of different extraction parameters (solvents, homogenization methods and different shaking systems to assist the extraction) and clean up strategies allowed the selection of

327	a method with recoveries between 70-120 % for endosulfan, its studied metabolites and
328	chlorpyrifos. Method 3 presents the best recoveries of the analytes as shown in Table 3.
329	The other methodologies evaluated did not accomplish DG-SANCO guidelines criteria
330	for pesticide residue analysis [16], obtaining recoveries under our laboratory conditions
331	<70 % or >200 % with RSDs between 1 and 10 %.
332	Table 3.
333	
334	3.1.2. Laboratory scale biomix (Matrix B)
335	As for matrix A, different methods were tested. This matrix required six method
336	comparisons due to its increased chemical complexity as soil, peat and bran were
337	together in the biomix (Table 4).
338	Method 5 presented the best results probably due to the formation of a biphasic
339	system through water addition and ultrasound assisted extraction which also made the
340	extraction time shorter. Also, ethyl acetate proved to have better extractability of the
341	analytes than acetonitrile (used in Methods 1 and 3). Traditional methods such as
342	soxhlet (Method 6) provide low recoveries which would lead to a overestimation of the
343	bioconversion. Performing a second extraction (Method 1) gave a large load of co-
344	extractives and unacceptable recoveries above 120 %.
345	G
346	Table 4.
347	
348	3.2. Method validation.
349	
350	3.2.1. Matrix A

351	The selected method yielded good results for endosulfan, its studied metabolites
352	and chlorpyrifos (Table 5); recoveries were between 72-109 % at all concentration
353	levels studied. The method was precise: its repeatability (RSDr) was below 11% and its
354	within-laboratory reproducibility (RSDwR) is below 17 % for all analytes at all the
355	evaluated levels, according to DG-SANCO guidelines criteria for pesticide residue
356	analysis [16]. The determined LOQs and LODs were adequate for the purpose of the
357	method: the quantification of chlorpyrifos and endosulfán, residues as well as its
358	metabolites in a fungi growing in cereal bran in semisolid media. The values for LOQs
359	and LODs are 1 mg kg ⁻¹ and 0.3 mg kg ⁻¹ respectively. It was not possible to perform the
360	sub sampling and sample comminution steps prior to solvent addition in order to
361	achieve good accessibility of the solvent to the analytes because the fungi grows
362	heterogeneously.
363	
364	Table 5.
365	
366	The fungal mycelium and the components of bran are responsible for marked
367	matrix effects when the amount of matrix injected is higher, being very important for
368	experiments were the initial endosulfan concentration was 1 mg/kg and at 25 mg kg ⁻¹

(Table 6). Matrix-matched calibration curves were used for quantitation at all concentration levels because the influence of the amount of matrix on the analyte response was very heterogeneous. The recommendation is to perform the quantitation step using matrix matched calibration curves when assessing pesticide dissipation and metabolites generation with a blank extract prepared in the same way as the samples.

- 374
- 375

Table 6.

One of the advantages of this method is that as the sample contained (fungi + semisolid culture medium + bran) is completely extracted, no sub sampling is necessary, so all problems related to representativeness are avoided. Also this characteristic improves the results in terms of repeatability and reproducibility.

381

382 3.2.2. Matrix B

The results obtained with the best method for chlorpyrifos, endosulfan, and the 383 different metabolites assayed are presented in Table 7. Recoveries were between 71-110 384 % at all concentration levels studied. The method was precise: its repeatability (RSDr) 385 was below 17 % and its within-laboratory reproducibility (RSDwR) is below 18 % for 386 all analytes at all the evaluated levels, according to DG-SANCO guidelines criteria for 387 388 pesticide residue analysis [16]. The determined LOQs and LODs were adequate for the purpose of the method: the quantification of endosulfan, chlorpyrifos residues and its 389 metabolites for a fungi growing in a laboratory biobed. The values for LOQs and LODs 390 are 1 mg kg⁻¹ and 0.3 mg kg⁻¹ respectively. 391

392

393

Table 7.

394

Matrix effects vary with the different analytes (Table 8), for example, endosulfan sulfate showed the highest matrix effect at all concentrations but for endosulfan ether it was zero. The recommendation is to always quantify with matrix matched calibration curves when assessing endosulfan, its metabolites or chlorpyrifos dissipation. These curves are generated with a blank extract prepared in the same way as the samples injecting the same amount of matrix co-extractives.

401

Table 8.

403

402

Ultrasound assisted extraction allowed better solvent accessibility to the analytes. The method has few steps and is easy to perform in the laboratory considering the complexity of the matrix composed of a microbiological organism grown in a biobed degrading recalcitrant pesticides.

A global strategy to study bioremediation of contaminants at laboratory scale, based 408 on the precise knowledge of the concentration of pesticides present in the biobed is 409 presented. Two very employed; semipersistent to persistent pesticides were used as 410 models. The extraction and clean up methodologies used for the sample preparation of 411 matrix A and B are adequate for the extraction of chlorpyrifos, endosulfan and its 412 413 metabolites as it was shown during the validation steps. These methods allow the study of the proposed bioreactor and provide tools for studying biotransformation processes of 414 415 endosulfan, its metabolites and chlorpyrifos by GC-ECD. Exploratory trials of 416 endosulfan and chlorpyrifos dissipation at laboratory scale biobeds yielded 87 % for endosulfan and 79% for chlorpyrifos degaradation after 27 days. Figure 1 shows the 417 chromatogram of the residual pesticides in the biobed after 9 and 27 days of 418 bioconversion respectively, proving the viability of the proposed bioreactor to degrade 419 the agrochemicals under study. 420

Further work is in progress seeking to optimize the conditions of the bioreactor. The results obtained are a step forward in the search of an environmentally friendly tool to diminish the impact of recalcitrant compounds in affected areas.

424

425 4. Acknowledgements

- 426 ANII_PR_FMV_2009_1_2942.
- 427 ANII_PR_FMV_2_2011_1_6740.
- 428 PEDECIBA Química.
- 429 LATU.
- 430 FAO/IAEA through RALACA
- 431
- 432 5. References
- 433
- 434 [1] A. Carter, How pesticides get into water and proposed reduction measures,
- 435 Pesticide Outlook 11 (2000) 149-156.
- 436 [2] E. Karanasios, N. Tsiropoulos, D. Karpouzas, On-farm biopurification systems for
- 437 the depuration of pesticide wastewaters: recent biotechnological advances and future
- 438 perspectives, Biodegradation 23 (2012) 787-802.
- 439 [3] P.J. Mason, I.D.L. Foster, A.D. Carter, A. Walker, S. Higginbotham, R.L. Jones,
- 440 I.A.J. Hardy, Relative importance of point source contamination of surface
- 441 waters: river Cherwell catchment monitoring study, in: Symposium on Pesticide442 Chemistry, Cremona, Italy, 1999.
- [4] J.N. Kreuger, E., Catchment scale risk-mitigation experiences key issues for
 reducing pesticide transport to surface waters, in: Pesticide behaviour in soils and water.
 Proceedings of a Symposium organized by the British Crop Protection Council,
 Brighton, UK, 2001.
- [5] T. de wilde, J. Mertens, P. Spanoghe, J. Ryckeboer, P. Jaeken, D. Springael,
 Sorption kinetics and its effects on retention and leaching, Chemosphere 72 (2008) 509-
- 449 516.

- [6] W.A.I. Pedro, J. Alvarez, Bioremediation and Natural Attenuation: Process
 Fundamentals and Mathematical Models, in: J.W. Sons (Ed.) A groundbreaking text and
 professional resource on natural attenuation technology, 2006, pp. 609.
- [7] M. Marinozzi, L. Coppola, E. Monaci, D. Karpouzas, E. Papadopoulou, U.
 Menkissoglu-Spiroudi, C. Vischetti, The dissipation of three fungicides in a biobed
 organic substrate and their impact on the structure and activity of the microbial
 community, Environ. Sci. Pollut. Res. 20 (2013) 2546-2555.
- [8] E. Karanasios, N.G. Tsiropoulos, D.G. Karpouzas, C. Ehaliotis, Degradation and
 Adsorption of Pesticides in Compost-Based Biomixtures as Potential Substrates for
 Biobeds in Southern Europe, J. Agric. Food Chem. 58 (2010) 9147-9156.
- 460 [9] S.B. Pointing, Feasibility of bioremediation by white-rot fungi. Applied
 461 Microbiology and Biotechnology 57 (2001) 20-33.
- [10] Č. Novotný, K. Svobodová, P. Erbanová, T. Cajthaml, A. Kasinath, E. Lang, V.
 Šašek, Ligninolytic fungi in bioremediation: extracellular enzyme production and
 degradation rate, Soil Biology and Biochemistry 36 (2004) 1545-1551.
- 465 [11] G. Davila-Vazquez, R. Tinoco, M.A. Pickard, R. Vazquez-Duhalt, Transformation
- 466 of halogenated pesticides by versatile peroxidase from Bjerkandera adusta, Enzyme and
- 467 Microbial Technology 36 (2005) 223-231.
- 468 [12] M.d.P. Castillo, L. Torstensson, Effect of biobed composition, moisture, and
 469 temperature on the degradation of pesticides, J. Agric. Food Chem., 55 (2007) 5725.
- 470 [13] A. Rivero, S. Niell, V. Cesio, M.P. Cerdeiras, H. Heinzen, Analytical methodology
- 471 for the study of endosulfan bioremediation under controlled conditions with white rot
- 472 fungi, J. Chromatogr. B, 907 (2012) 168-172.
- 473 [14] J. A. Bumpus, S. N. Kakar, R. Coleman, Fungal degradation of organophosphorous
- 474 insecticides, Applied biochemistry and biotechnology 39 (1993) 715-726.

- [15] European Food Safety Authority (EFSA). Conclusion on the peer review of the
 pesticide human health risk assessment of the active substance chlorpyrifos. EFSA
 Journal 12 (2014) 3640.
- 478 [16] European Commission Health & Consumer Protection Directorate-General.
- 479 Guidance document on analytical quality control and validation procedures for pesticide
- residues analysis in food and feed. In SANCO/12571/2013, 2013.
- 481 [17] E. Karanasios, N.G. Tsiropoulos, D.G. Karpouzas, U. Menkissoglu-Spiroudi,
- 482 Novel biomixtures based on local Mediterranean lignocellulosic materials: Evaluation
- for use in biobed systems, Chemosphere 80 (2010) 914-921.
- 484 [18] E.C. Karanasios, N.G. Tsiropoulos, D.G. Karpouzas, Quantitative and qualitative
- differences in the metabolism of pesticides in biobed substrates and soil, Chemosphere
 93 (2013) 20-28.
- [19] M. Omirou, P. Dalias, C. Costa, C. Papastefanou, A. Dados, C. Ehaliotis, D.
 Karpouzas, Exploring the potential of biobeds for the depuration of pesticidecontaminated wastewaters from the citrus production chain: Laboratory, column and
 field studies, Environmental Pollution 166 (2012) 31-39.
- [20] L. Coppola, M.d.P. Castillo, E. Monaci, C. Vischetti, Adaptation of the biobed
 composition for chlorpyrifos degradation to southern Europe conditions, Journal of
 Agricultural and Food Chemistry 55 (2007) 396-401.
- 494

	ACCEPTED MANUSCRIPT
495	
496	
497	Captions for Figures and Tables
498	Table 1. Extraction and clean up methods comparison for matrix A.
499	Table 2. Extraction methods comparison for matrix B.
500	Table 3. Recoveries (Rec (%)) comparison for the evaluated methods to determine
501	endosulfan and chlorpyrifos and their respective RSDs (%) in matrix A.
502 503	Table 4. Recoveries (%) and their respective RSDs (%) for the 6 evaluated methods in matrix B.
504 505	Table 5. Recoveries (%), repeatability (RSDs %) and within-laboratory reproducibility (RSDwR %) of the evaluated pesticides in matrix A.
506	Table 6. Matrix effect for chlorpyrifos, endosulfan and its metabolites at 1 and 25 (mg
507	kg ⁻¹) by GC-ECD in matrix A.
508	Table 7. Recoveries (%) and RSDs (%) of the evaluated pesticides for the analytical
509	method developed in matrix B; levels 1, 25 and 50 mg kg ⁻¹ for endosulfan and
510	metabolites and levels 1, 25 and 60 mg kg ⁻¹ for chlorpyrifos
511	Table 8. Matrix effects for endosulfan, chlorpyrifos and its metabolites at 1, 25, 50 mg
512	kg ⁻¹ by GC-ECD in matrix B. 1, 25 and 50 mg kg ⁻¹ for endosulfan and it metabolites
513	and 1, 25 and 60 mg kg ⁻¹ for chlorpyrifos.
514	Figures:

- 515 Figure 1. GC-ECD chromatogram of the chlorpyrifos and endosulfan residues in labs
- biobeds after 9 and 27 days, showing their degradation. Black trace: residues after 9
 days; purple trace: residues after 27 days.

	ACCEPTED MANUSCRIPT								
519									
520	Table 1.								
521									
522		Metho	od 1	Method 2	Method	3			
523	Sample (g)	15		15	15				
524	Solvent, volume (mL)	AcOE	t, 40	AcOEt, 40	MeCN,	15			
525	Water addition (mL)	-		-	15				
526	Agitation/time (h)	Stoma	cher/0.08	Shaker/24	Shaker/	2			
527	Clean up: NaSO ₄ /Silica colu	ımn (g)	20 / 30	20/30	-	X			
528	Elution: solvent, volume (m	L)	AcOEt, 100	AcOEt, 100	-	9			
529	Salting out: NaCl/MgSO4 (g		-	-	2/20				
530	Agitation/time(h)		-	-	Shaker/	5			
531	Centrifugation (min)/(rpm)		-	-		10/3000			
532	Final aliquot (mL)		0.1	0.1	0.5				
533			7 6.						
534			O						
	0	2							
	G								

Table 2.

ACCEPTED MANUSCRIPT

537		Method1	Method2	Method3	Method4 Met	hod5 M	ethod 6
538	Sample (g)	15	15	15	10 10		10
539	Humidity (%)	40	40	40	70 70		40
540	Solvent (mL)	MeCN(30)	AcOEt(30)	MeCN(30)	AcOEt(50)	AcOE	t(50) AcOEt(250)
541	Water (mL)	15	-	-	*	-	-
542	Extraction	shaker	ultras	ultras	shaker	ultras	soxhlet
543	Time (h)	2	0.15	0.15	5	0.45**	2.30
544	MgSO ₄ /NaCl	20(g)/2(g)	-	-	-	9	-
545	Solvent (mL)	MeCN(30)	-	-	- 6	-	AcOEt(50)
546	Shake (h)	5	-	-	5	-	-
547	Centrifugation	n (min/rpm)	10/3000	-	10/3000	10/300	00 -
548	Aliquot (mL)	0.5	0.5	0.5	0.5	0.5	0.5
549							
550							
		CO					
		G					
	P						

				ACCE	PTED	MAN	USCE					
551	Table 3.			TOGE								
552												
553		Endos	sulfan o	X		Endos	ulfan β	}		Chlo	rpyrifos	
554	Rec	RSDr	RSD	wR	Rec	RSDr	RSDv	wR	Rec	RSD	r RSDwl	R
555		(%)	(%)	(%)		(%)	(%)	(%)		(%)		(%) (
556	Method 1	41	3	6		43	4	9	201	20		10
557	Method 2	63	2	3		65	1	1	205	12	6	
558	Method 3	83	3	4		91	3	4	76	8	10	
559												
560										9		
561									C			
562								, C				
563												
564							0					
565					~							
- 66					0							
567												
569			.0									
570		G										
571												
572												
573												
574												
575												
576												
5/7												
	25											

Table 4.

581			Endos	ulfan α			Endos	ulfan β		Chlor	oyrifos		
582		Rec	RSDr	RSDw	'R	Rec	RSDr	RSDw	'R	Rec	RSDr	RSDw	vR
583		(%)		(%)	(%)	(%)		(%)	(%)	(%)		(%) (%)
584	Method 1	170		-	-	169		-	-	220		-	-
585	Method 2	111		1	1	115		2	1	172		0	0
586	Method 3	55		1	2	62		1	2	60		2	3
587	Method 4	122		8	7	119		10	8	116		8	7
588	Method 5	94		10	11	92		10	11	75		5	7
589	Method 6	67		10	15	65		10	16	60		8	13
590													
591							0						
					2								
			0										
		G											

		A	CCEPTED	MANUSCF	RIPT		
592	Table 5.						
593							
594		1 (mg	; kg ⁻¹)		25 (mg kg ⁻¹)		
595	Rec(%	6)	RSDr(%)	RSDwR(%)	Rec(%)	RSDr(%)	RSDwR(%)
596	Endosulfan ether	83	3	3	86	3	8
597	Endosulfan alcohol	78	11	16	106	8	14
598	Endosulfan α	80	3	11	89	5	10
599	Endosulfan β	80	4	14	96	7	8
600	Endosulfan sulfate	90	3	12	81	1	17
601	Chlorpyrifos	72	5	13	109	7 9	
602					G		
603					9		
604							
605							
606							
607			0	>			
608							
609		.0					
610	C						
612							
613							
614							
615							
616							
617							

Accepted manuscript

	ACCEI	PTED MANUSCRIPT
Table 6.		
	1 (mg kg ⁻¹)	25 (mg kg ⁻¹)
	(%)	(%)
Endosulfan ether	-49	11
Endosulfan alcohol	36	52
Endosulfan α	-39	21
Endosulfan β	-34	-1.6
Endosulfan sulfate	27	49
Chlorpyrifos	-7	25
		5
		2
		6
	*	
	0	
	.0	
G		

ACCEPTED MANUSCRIPT												
632												
633	Table 7.											
634												
635		1(mg	kg ⁻¹)		25(mg	; kg ⁻¹)		50/60(mg kg	¹)		
636		Rec	RSDr	RSDw	R	Rec	RSDr	RSDw	/R	Rec	RSDr	RSDwR
637		(%)	(%)	(%)		(%)	(%)	(%)	(%) (%	6)(%)		
638	Endosulfan ether	76	5	11	81	15	13	83		17	18	
639	Endosulfan alcohol	71	8	7	90	13	17	110		15	15	
640	Endosulfan a	80	6	10	84	13	12	94		15	16	
641	Endosulfan β	77	5	17	87	14	12	101	\mathbf{O}	18	16	
642	Endosulfan sulfate	90	9	11	97	13	10	106		15	16	
643	Chlorpyrifos	102	8	7	110		15	16	105	12	10	
644												
645						b						
646					\mathbf{O}							
647				Ó	*							
648												
649			2									
650	_	5										
651												
652												
653												
654												
655												
656												
657												

	ACCEPTED MANUSCRIPT									
658										
659										
660	Table 8.									
661										
662		1(mg kg ⁻¹)	25(mg kg ⁻¹)	50/60(mg kg ⁻¹)						
663		(%)	(%)	(%)						
664	Endosulfan ether	0	0	0						
665	Endosulfan alcohol	56	28	24						
666	Endosulfan α	55	3	12						
667	Endosulfan β	21	2	7						
668	Endosulfan sulfate	86	53	77						
669	Chlorpyrifos	31	29	46						
670				N.						
671				3						
			. 0							
			Ó.							
		X								
		2								
	NG.									



